Long-Term Trend of Indoor Volatile Organic Compounds – a 15-Year Follow-Up Considering Real Living Conditions

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Key Words
Environmental medicine · Hygiene · Indoor exposure · Long-term trend · VOCs

Abstract
Indoor volatile organic compounds (VOCs) may be associated with adverse health effects. The aim of this study was to investigate whether the indoor VOCs exposure burden has undergone changes during a past 15-year period, exemplarily in a German city. The study was limited to homes with families. To avoid possible regional-specific confounder effects, this assessment was conducted for this urban centre with a sufficiently long and continuous time series (1994–2008).

VOCs were determined using passive sampling over 4 weeks followed by liquid extraction and analysis with gas chromatography-mass spectrometry. All measurement periods were equally distributed over the seasons and years.

A substantial reduction in the total sum of VOCs and aromatics could be demonstrated as well as a trend of decrease concentrations for alkanes. No such changes could be observed for cycloalkanes, chlorinated hydrocarbons and terpenes group.

Possible reasons for the reduction of indoor VOCs exposure are most likely due to the applied regulations/guidelines for the VOCs reduction in indoor-relevant products. Problems are, however, regarding the slight increase in terpene exposure. These chemicals have been found to be associated with quite specific health effects. Therefore, future emphasis should be directed not only towards the general reduction of anthropogenically generated VOC exposures in the outdoor environment but also on the reduction of health-relevant VOCs, especially indoors.

Introduction
During the last 20 years, epidemiologic investigations have shown associations between indoor exposure to volatile organic compounds (VOCs) and signs and symptoms of ill health [1–8]. Even cell experiments have
demonstrated effects of VOCs on cellular processes showing (pre)-pathologic propensities [9]. For this reason, VOCs have incurred a special importance in the regulation of indoor air quality.

The WHO, the EU as well as individual countries have already initiated steps to reduce the VOCs burden in populations (see among others References [10–13]).

Over the years, the associations of VOCs to ill health have been controversially discussed [14,15]; however, data are accumulating, showing that even small dosages may already account for these observed health effects [2,4,8]. The question is, therefore, has the VOCs indoor burden changed in any way, with regards to its concentration as well as in the proportion of its components compared to the total exposure. It is possible that any changes in the VOCs component profile could influence, in terms of positive or negative trends, the health relevance of the mixture. Furthermore, changes observed over longer periods of time indicate that components, possibly originating from products associated with the indoor environment, can be removed and replaced by others in the indoor environment. This may have implications for Public Health services, which would require surveillance programmes or to request product changes for building applications.

This paper presents the evaluation based on the VOCsum concentration. The VOCsum index is ideal for diverse regulatory sectors, especially as a general indicator for total VOCs in enclosed areas. The main importance of the index though is that it is derived from the average (and obviously reached) concentrations, exceedance of the index value would reveal evidence of an existing source.

However, this is critical in regards to the sampling and the analytical techniques. According to the European Working Group [16], the definition of the VOCs sum relates to those VOCs that are detected within one chromatographical window. These concepts are definitely applied sensibly in intervention studies or quality controls for product emissions. If VOCs are selectively measured and summed up, the proportion of individual constituents, contributing towards the VOCsum total concentration, could still vary. This is critical for the assessment of indoor environments, in case that the proportion of health relevant chemicals has increased over time.

On the one hand the assessment of the VOCsum is problematic. Those substances which are difficult to sample with conventional sampling techniques like reactive organic compounds, e.g., aldehydes or intermediary reaction products, are not covered. On the other hand long-term trend studies are useful in the assessment of the indoor VOCs burden to produce an overview about the time trend of a single VOC which contribute predominantly to VOCsum. This fact is of importance especially for those components that are covered in routine analyses being carried out by local health departments. To reveal contributions of indoor VOCs to health disorders, investigations were directed to assess indoor air exposure by means of measurement of a standard mixture of VOCs. This standard mixture is based on a mixture proposed by Molhave, Bach, and Pedersen [17] to control the indoor air environment. It has to be assumed that a single VOC as a part of this mixture can change their proportion over several years.

Considering these facts, the aim of this study was to find out whether the absolute concentration of different VOC groups and the relative ratio between VOC groups have undergone changes over a past 15-year period.

It should be pointed out explicitly that the presented paper only deals with the description of the VOCs exposure and not with resulting health effects. Because of the complexity it is necessary to deal with the health association in a separate paper.

**Materials and Methods**

**Design**

As a part of several epidemiologic studies (studies described for example in References [18–20]), among other investigations, VOCs were measured in the participants' residences in order to clarify association issues between exposure and ill health.

The studies were approved by the ethics committee of the University of Leipzig.

For the here presented analyses, only measurements were considered which pertain to the study population and their residences in the city of Leipzig. There were 4725 participants involved in the study. From this population in total, 2679 individual measurements were performed over a time span of 15 years, starting 1994 to 2008.

The basis for the here presented data are these different longitudinal epidemiological cohort studies starting with the participants births. As mentioned, one of the aims of these studies was to find out whether the indoor environment can contribute to “negative” health outcomes. Within the spectra of the investigated indoor load, VOCs were determined.

The participants respectively the parents of the participating children received a passive sampler, which was exposed in the room of the apartment/house in which the
The participant spent most of their time during a day. As a rule, this room should be the children’s bedroom or the living room of the family (the parents provided the information regarding the room in the epidemiological questionnaire). The passive sampling system was exposed at the first time around the birth and in the following years (with interruptions – not every year) around the birthday of the children. The children were recruited at the beginning of their birth. The time of recruitment was, dependent on the study, 1 or 2 years. Since the birthdays are equally distributed over the year, the measurement times were also equally distributed over the year. That means, no season would be over-represented. Furthermore, since the exposure time was 4 weeks, peak loads would not be of significance.

The number of samples per home was different and depended on the participation duration in the study. As a rule, one to three samples per apartment/house were obtained during any sampling occasion over the investigation period.

Table 1 presents the number of measurements taken in each corresponding year. These indoor measurements were taken randomly among the participating families. Since the number of study participants was different for the different studies and since in some periods two studies have been overlapped, the total sample size per year was different.

### Sampling

All measurements were taken using passive samplers [21]. The passive sampling method has the advantage to record a mean load over a sampling period of 28 days and from that point of view this method avoids short-time peaks (spikes). This is an advantage since short-time measurements would assess the VOC exposure in a very small time window and are therefore not as representative as long-term passive measurements. The applicability of passive air sampling for VOCs assessment has been validated sufficiently in the literature [22–26].

The passive samplers were deposited for 4 weeks in the participants’ residences in those rooms in which the family spent most of its time. As a rule, one sampler per family home was installed.

### Analysis

Briefly, the VOC-coated active charcoal layer of the 3 M passive sampler was extracted with 1.5 mL carbon disulfide (for details see References [25,26,28]). The 30-min extraction time was followed by a quantitative VOC analysis using the capillary gas chromatography (Autosystem-Gas Chromatography, Perkin Elmer, Flame ionisation detector (FID)/Electron capture detector (ECD)).

- capillary columns Polyphenylmethylsiloxane (PVMS/54), Dimethylsiloxane (DMS), 50 m, ID (internal diameter) = 0.32 mm, layer thickness 1 m
- carrier gas nitrogen, 14.5 psi
- injection volume 2 L, split 1:10
- temperature 40°C (program 2 K min\(^{-1}\) to 150°C)

A five-point calibration function was determined for several identified components based on the integrated FID.

<table>
<thead>
<tr>
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<th></th>
<th></th>
<th></th>
<th></th>
</tr>
</thead>
<tbody>
<tr>
<td>N</td>
<td>88</td>
<td>266</td>
<td>158</td>
<td>286</td>
<td>287</td>
<td>100</td>
<td>556</td>
<td>377</td>
<td>312</td>
<td>249</td>
</tr>
</tbody>
</table>

N: As a rule every number stands for one home.
signal area. This calibration function was used to determine the VOC mass adsorbed on the charcoal layer. The average concentration over the sampling interval was then calculated for each component (in $\text{mg} \cdot \text{C}_0 / \text{s}^3$) in accordance with equation (1) adopted from the 3 M Application Bulletin [21].

$$c = \frac{m \cdot A}{t \cdot r}$$

where $m$ = the absolute mass amount of the adsorbed component (in $\mu$g), $t$ = the sampling interval (in min), $r$ = the recovery coefficient and $A$ = a constant.

“$A$” includes the diffusion coefficient of the component, the diffusion area of the charcoal pad and the diffusion distance inside the sampler. For some components, the constant $A$ was adopted from a list published in the 3 M Application Bulletin [21]. For any not listed compound, the diffusion constants were calculated according to a procedure proposed by Pannwitz [29].

**Table 2.** Volatile organic compounds measured in all measurement campaigns

<table>
<thead>
<tr>
<th>Alkanes</th>
<th>Heptane, octane, nonane, decane, undecane, dodecane, tridecane</th>
</tr>
</thead>
<tbody>
<tr>
<td>Cycloalkanes</td>
<td>Methylcyclopentane, cyclohexane, methylocyclohexane</td>
</tr>
<tr>
<td>Aromatics</td>
<td>Benzene, toluene, ethylbenzene, m,p-, o-xylene, styrene, 2-, 3-, 4-ethyltoluene</td>
</tr>
<tr>
<td>Chlorinated hydrocarbons</td>
<td>Chlorobenzene, trichlorethylene, tetrachlorethylene</td>
</tr>
<tr>
<td>Terpenes</td>
<td>$\alpha$- and $\beta$-pinene, 2- and 3-carene, limonene</td>
</tr>
</tbody>
</table>

**Table 3.** Sum concentrations of all VOCs (VOCsum) ($\mu$g m$^{-3}$)

<table>
<thead>
<tr>
<th>Year</th>
<th>N</th>
<th>Mean</th>
<th>Median</th>
<th>Geometric mean</th>
<th>90% percentile</th>
<th>95% percentile</th>
<th>Maximum</th>
<th>Standard deviation</th>
<th>Standard error</th>
</tr>
</thead>
<tbody>
<tr>
<td>1994</td>
<td>88</td>
<td>287.2</td>
<td>221.2</td>
<td>229.3</td>
<td>571.1</td>
<td>744.5</td>
<td>1352.7</td>
<td>214.4</td>
<td>22.9</td>
</tr>
<tr>
<td>1995</td>
<td>266</td>
<td>272.2</td>
<td>180.0</td>
<td>165.5</td>
<td>495.1</td>
<td>739.3</td>
<td>4597.1</td>
<td>416.7</td>
<td>25.5</td>
</tr>
<tr>
<td>1997</td>
<td>158</td>
<td>190.6</td>
<td>98.7</td>
<td>105.8</td>
<td>411.2</td>
<td>716.2</td>
<td>2712.1</td>
<td>313.6</td>
<td>24.9</td>
</tr>
<tr>
<td>1998</td>
<td>286</td>
<td>179.5</td>
<td>136.7</td>
<td>127.1</td>
<td>354.7</td>
<td>497.7</td>
<td>1923.8</td>
<td>178.1</td>
<td>10.5</td>
</tr>
<tr>
<td>1999</td>
<td>287</td>
<td>179.8</td>
<td>142.5</td>
<td>141.0</td>
<td>366.2</td>
<td>444.6</td>
<td>940.5</td>
<td>134.8</td>
<td>8.0</td>
</tr>
<tr>
<td>2000</td>
<td>100</td>
<td>203.1</td>
<td>146.4</td>
<td>150.2</td>
<td>418.2</td>
<td>540.6</td>
<td>1505.8</td>
<td>196.0</td>
<td>19.6</td>
</tr>
<tr>
<td>2001</td>
<td>556</td>
<td>181.5</td>
<td>144.8</td>
<td>135.2</td>
<td>364.2</td>
<td>465.8</td>
<td>1136.3</td>
<td>143.6</td>
<td>6.1</td>
</tr>
<tr>
<td>2004</td>
<td>377</td>
<td>170.6</td>
<td>126.1</td>
<td>120.8</td>
<td>323.1</td>
<td>453.7</td>
<td>1556.6</td>
<td>165.0</td>
<td>8.5</td>
</tr>
<tr>
<td>2007</td>
<td>312</td>
<td>141.4</td>
<td>82.6</td>
<td>91.9</td>
<td>299.0</td>
<td>406.4</td>
<td>1848.7</td>
<td>176.4</td>
<td>10.0</td>
</tr>
<tr>
<td>2008</td>
<td>249</td>
<td>123.5</td>
<td>71.1</td>
<td>72.0</td>
<td>299.3</td>
<td>397.1</td>
<td>1570.4</td>
<td>157.6</td>
<td>10.0</td>
</tr>
</tbody>
</table>

VOCs: volatile organic compounds.

**Fig. 1.** Temporal course of the VOCsum – concentration 1994–2008 (+/−95% CI SEM) (SEM – standard error of mean).
The detection limit of the investigated components was between 0.1 and 1 µg m⁻³.

Depending on the respective measurement campaign, a varied number of VOC components were measured. At present, a total of 87 individual components can definitely be determined. For the here presented assessment and for reason of comparability, only those 26 components are included in the analyses, which were measured in every campaign and for which a complete set of data was available (Table 2).

**Statistical Analyses**

The statistical analyses were conducted using Statistica [30] and Excel [31]. Descriptive statistics are presented as (arithmetic) means and standard deviations and absolute numbers (sample sizes) and (relative) frequencies.

The trend was calculated using the Mann-Kendall-test. This test is a non-parametric test and examines the signs of pair-wise differences of the VOC time series.

**Table 4.** The p-values for the temporal trend of the individual VOC components (bold: significant p < 0.05; bold in italics: trend significant p < 0.10; Mann-Kendall trend test)

<table>
<thead>
<tr>
<th>Alkanes</th>
<th>p</th>
<th>Cycloalkanes</th>
<th>p</th>
<th>Terpenes</th>
<th>p</th>
<th>Aromatics</th>
<th>p</th>
</tr>
</thead>
<tbody>
<tr>
<td>Hexane</td>
<td>0.22</td>
<td>Methylcyclopentane</td>
<td>0.58</td>
<td>a-pinene</td>
<td>0.79</td>
<td>Benzene</td>
<td>0.06</td>
</tr>
<tr>
<td>Heptane</td>
<td>0.31</td>
<td>Cyclohexane</td>
<td>0.31</td>
<td>b-pinene</td>
<td>0.32</td>
<td>Toluene</td>
<td>0.02</td>
</tr>
<tr>
<td>Octane</td>
<td>0.06</td>
<td>Methylcyclohexane</td>
<td>0.31</td>
<td>2-carene</td>
<td>0.46</td>
<td>Ethylbenzene</td>
<td>0.09</td>
</tr>
<tr>
<td>Nonane</td>
<td>0.06</td>
<td></td>
<td></td>
<td>3-carene</td>
<td>0.69</td>
<td>m + p-xylene</td>
<td>0.06</td>
</tr>
<tr>
<td>Decane</td>
<td>0.08</td>
<td>Chlorinated HC</td>
<td></td>
<td>limonene</td>
<td>0.11</td>
<td>α-xylene</td>
<td>0.03</td>
</tr>
<tr>
<td>Undecane</td>
<td>0.06</td>
<td>Chlorobenzene</td>
<td>0.54</td>
<td></td>
<td></td>
<td>Styrene</td>
<td>0.20</td>
</tr>
<tr>
<td>Dodecane</td>
<td>0.16</td>
<td>Trichlorethylene</td>
<td>0.06</td>
<td></td>
<td></td>
<td>2-ethyltoluene</td>
<td>0.16</td>
</tr>
<tr>
<td>Tridecane</td>
<td>0.11</td>
<td>Tetrachlorethylene</td>
<td>0.03</td>
<td></td>
<td></td>
<td>3-ethyltoluene</td>
<td>0.03</td>
</tr>
<tr>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td>4-ethyltoluene</td>
<td>0.13</td>
</tr>
</tbody>
</table>

VOCs: volatile organic compounds; HC: hydrocarbons.
Results

Concentration Trends (Absolute)

Using the measured concentration data obtained in the different years, the sum concentrations (VOCsum) of all VOCs (listed in Table 2) were calculated and are presented in Table 3. The bases for the calculations are the detected sum concentrations of each individual residence for each measurement year. Figure 1 depicts the temporal course of the VOCsum concentration during the years 1994 to 2008.

The time trend was tested using the Mann-Kendall trend test. A significant trend seems to exist if the $p < 0.05$. The Mann-Kendall trend test shows the temporal changes of the decrease for all VOCs (VOCsum) ($p = 0.05$) and the group of aromatics ($p = 0.02$), as well as a demonstrable trend of a reduction for the group of alkanes ($p = 0.08$). The group of cycloalkanes, chlorinated hydrocarbons and terpenes do not indicate any time-dependent trend change (Figure 2). Figure 2 shows the temporal course of those listed VOC groups presented in Table 2. Table 4 shows the $p$-values of the individual VOCs.

Additional to the trend of VOCsum within the groups of VOCs such as alkanes, aromatics etc., the individual substances were differentiated to show contribution to the respective chemical groups. Table 4 shows for which single compounds within the groups of VOCs, a time trend can be observed. The trend was calculated using the Mann-Kendall trend test. If the $p$-value is lower than 0.05, it can be assumed that the single VOC would show a time trend which is significant over the years.

Trends of the Concentrations of the Proportional Compounds (Relative)

With a generally decreasing exposure burden in the temporal course, the proportion of individual substances on the total burden shows a clear shift. While the proportions of the alkanes and aromatics appear to decrease, the terpenes’ proportion on this burden showed an increasing trend (Figure 3).

Assuming simplistically a linear function course, the decrease of the alkane group would amount to about $-4.1 \mu g m^{-3}$ per year, for the cycloalkanes $-0.4 \mu g m^{-3}$, the aromatics about $-4.7 \mu g m^{-3}$ and the chlorinated hydrocarbons around $-0.5 \mu g m^{-3}$ per year, whereas an increase of $+0.3 \mu g m^{-3}$ per year was found for the terpenes. Assuming a significance of $p = 0.05$ with a correlation coefficient differing from 0, only the given trends for the alkanes and aromatics were found to be significant.

Discussion and Conclusion

Indoor-VOCs can play an important role in the assessment of potential health effects associated with an indoor air pollution burden. For this reason, it is essential to be aware of any changes in VOC profiles and to follow their trend analyses and assess the percentages of the
individual compounds. These steps would allow regulation, setting of standards and implementations of preventive measures.

This study indicates that over the past 15 years (1994 to 2008), clear trends could be observed towards changes in the portion of 26 VOCs under investigation within these chemical “mixtures.” Judging from that, one can assume that variations in the composition may derive different symptoms and also a difference in the prevalence of symptoms.

The presented results show a clear reduction in the sum of all VOCs, the alkanes and aromatics, while the burden of cyclo-alkanes and chlorinated hydrocarbons remained constant and that of the terpenes increased slightly (not significantly). In 2008, the sum concentration was only half of the 1994 concentration. This reduction was caused by a decrease in alkanes and aromatics concentrations. In fact, the sum reduction corresponded almost exactly with the balance change of the VOC sum. The reduction in the VOCsum concentration and the almost constant concentration of terpenes had led to a higher proportion of terpenes.

What are the potential reasons for these changes especially of the decreasing trend of VOCsum respectively for the VOC mixtures?

- All in all, contributing clearly to the reduction is the growing utilization of indoor materials with low VOCs content. Beginning in the mid-1990, solvents with a low boiling point generally used in the commercial/industrial production of lacquers and paints, were replaced by glycols and higher alkanes, which do not fall under the definition of solvents (boiling point <200°C). What we see clearly is a decline of “classical” solvents. Possibly also due to other regulations, recommendations and voluntary limitations by the industry have also contributed to the observed effects.

The appearance of other chemical substances in indoor air exposure, e.g., long chain alkanes, branched alkanes or glycols, was not investigated. There are hints in the integrated chromatograms of higher amounts of C_{14}-C_{16} alkanes. This remains a speculation until a broad epidemiological study has been carried out.

- Taking the individual constituents into consideration, the greatest contribution stems from toluene and the higher alkanes. A regulation for toluene was established in Germany, in 1996 [32]. Following that in 1997, toluene showed an immediate reduction by 60 μg m^{-3} down to 23 μg m^{-3} as compared to 1995/96. The subsequent decrease of 8 μg m^{-3} (from 23 μg m^{-3} to 15 μg m^{-3}; the mean between 1999 and 2008) was much less spectacular.

- Moreover, it may be assumed that the additives to the solvents for indoor paints and lacquers including glues (e.g., floor coverings) have undergone changes in their composition.

- It is also possible that the various epidemiological studies, which have drawn the attention to the VOC-associated adverse health effects, and the subsequent information dissemination to the parents of the children who participated in these studies, contributed to a more careful handling in the use of VOC-containing products. The first of these information events for parents took place in 1996, following the partial analysis of the first study [18]

While “classical” solvents based on their high volatility contributed not longer than a few days or weeks to the indoor VOC exposure burden [28], products, labelled as solvent-free, which contain compounds with a higher boiling point (>200°C), can possibly emit these into the indoor air for months or even years. For this reason, critical inspection is required and the profile of the indoor exposure burden would need to be looked upon analytically and methodologically in a new and more detailed way. Moreover, based on these presented data, it is suggested that there is a definite need for standards to be set and processes to be developed, which enable an assessment of the building materials as well, in order to detect the sources of VOC emittants with health consequences.

Some limitations of the study are:

- It cannot be excluded that other than the measured VOCs could show a different trend. The selection was based on the requirement to be comparable with other studies that focussed on the same or similar selection of VOCs [25,26]. Furthermore, the complete time series could only be generated for the considered 26 VOCs;

- The different sample size per year during the study period. This may lead to different levels of significance but can be neglected since the sample size in all cases was higher than 100 except 1994 (see Table 1). The variation in the sample size was unavoidable since the number of measurements was dependent on different epidemiological studies with different designs;

- It is impossible to control the resulting bias regarding the influence of the dissemination of study results in
the past. A main aim of the epidemiological studies in the past was to communicate to the health offices and to the public in order to prevent allergies: Which activities should the parents avoid? It can be assumed that some parents applied the imparted recommendations. This may have affected the overall VOCs assessment in later years. Important information in this context is to avoid renovation activities during pregnancy and in the first years of life.

The main advantages of the presented investigations are:

- Compared with other studies, this study includes a very large number of measurements over a time period of 15 years. No time of year and/or type of apartment/house would be over represented due to the applied study design.
- The chosen measurement method – passive sampling over 4 weeks – ensures that activities and peak concentrations generated by short-time emissions (e.g., of smoking, life style events such as burning of incense, etc.) would not influence the results in a significant way (short-term influence vs. long-term average).

To what extent the results may be transferable to other regions should be addressed by further studies.

Acknowledgements

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